A stable temperature may favour continuous reproduction by *Theodoxus fluviatilis* and explain its high densities in some karstic springs

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ABSTRACT

A stable temperature may favour continuous reproduction by *Theodoxus fluviatilis* and explain its high densities in some karstic springs

*Theodoxus fluviatilis* is a common gastropod in many karstic springs in central Portugal. We investigated the possible reasons for the near-total restriction of this species to these springs. We first determined the spatial distribution of the species within a spring (Anços) and related the densities at sampling-patch scales to selected physical and chemical variables. We then determined the densities at several locations downstream from the spring and related these densities to selected physical and chemical variables. Finally, we assessed the population dynamics of the gastropod in the spring. In the spring, *T. fluviatilis* was more abundant in shallow areas with a rapid current and cobble-boulder substrates. In June-July 2006, the mean densities of *T. fluviatilis* in the spring varied from ∼ 10 to ∼ 9000 individuals m⁻² but decreased to zero 3800 m downstream. The physical and chemical changes along the stretch studied were minor; no significant correlations (Spearman rank correlation; *p* > 0.05) were observed between the gastropod abundances and the measured environmental variables or the PCA axes. In the spring, juveniles and egg masses were observed year-round, suggesting continuous reproduction. The temperature at the spring was fairly constant, ranging from 15.3 to 16.6 °C during the year (daily variation < 0.1 °C). Nine hundred metres farther downstream, the daily temperature variation reached 5 °C. We propose that the high densities of *T. fluviatilis* in the Anços spring are maintained by continuous recruitment mediated by the constant high temperature.

Key words: Secondary production, substrate, population dynamics, Portugal, macroinvertebrates.

RESUMEN

La estabilidad térmica en manantiales kársticos puede favorecer la reproducción continua de *Theodoxus fluviatilis*, explicando así sus altas densidades

*Theodoxus fluviatilis* es un gasterópodo común en muchos de los manantiales kársticos del centro de Portugal. La pregunta a la que intentamos responder es ¿por qué la especie está prácticamente restricta a los manantiales? Para ello determinamos primero la distribución espacial de *T. fluviatilis* en el manantial calcáreo del río de Anços y relacionamos la densidad a nivel de hábitats del tramo con algunos parámetros fisicoquímicos. Posteriormente determinamos densidades en varios puntos río abajo y las relacionamos con las condiciones químicas y físicas. Finalmente, determinamos la dinámica de poblaciones en el manantial. En el manantial kárstico *T. fluviatilis* fue más abundante en las zonas poco profundas con corriente y substrato pedregoso. En Junio-Julio de 2006, las densidades medias de *T. fluviatilis* variaron entre ∼ 10 y ∼ 9000 individuos m⁻² en el manantial, pero bajaron a cero 3800 m río abajo. Los cambios fisicoquímicos a los 3800 m fueron mínimos; no fueron observadas correlaciones significativas (correlación de rangos de Spearman, *p* > 0.05) entre la abundancia de *T. fluviatilis* y las variables ambientales o los ejes de Análisis de Componentes Principales. En el manantial, juveniles y masas de huevos fueron observadas durante todo el año y no se observaron cohortes, sugiriendo reproducción continua. La temperatura en el manantial varió entre 15.3 y 16.6 °C durante todo el año (variación diaria < 0.1 °C). Novecientos metros río abajo las variaciones térmicas diarias llegaron a los 5 °C. Proponemos pues que la elevada densidad de *T. fluviatilis* en el manantial kárstico de Anços se mantiene gracias a las temperaturas elevadas y constantes que favorecen su continua reproducción.

Palabras clave: Producción secundaria, sustratos, dinámica de poblaciones, Portugal, macroinvertebrados.
INTRODUCTION

Theodoxus fluviatilis (Linnaeus, 1758) is a freshwater gastropod occurring throughout Europe and Anatolia in lakes, streams, rivers, canals and even estuaries and brackish waters up to 60 m in depth (Fretter & Graham, 1978; Zettler et al., 2004; Bunje, 2005; Kirkegaard, 2006; Bunje & Lindberg, 2007). These gastropods are occasionally numerous, particularly in clean waters with a high calcium content (Carlsson, 2000), where they feed on biofilms, ingesting mainly diatoms, but also feed on blue-green algae and detritus (Fretter & Graham, 1978; Zettler et al., 2004; Kirkegaard, 2006). The life-history of T. fluviatilis is little known despite the species’ wide distribution. Reproduction, embryo development and egg hatching have been reported to occur at temperatures above 10-12°C (Fretter & Graham, 1978; Kirkegaard, 2006). Embryo development has been reported to last for 30 days at 25°C and for 65 days at 20°C (Kirkegaard, 2006). Temperature also affects development, survivorship and seasonal density variation (Kirkegaard, 2006). To our knowledge, all previous studies of this gastropod have been conducted in temperate seasonal systems. We observed that T. fluviatilis was abundant in and almost entirely restricted to many karstic springs in central Portugal. However, its densities decreased rapidly downstream. To our knowledge, the reason for this distribution in central Portugal and in other locations has not yet been addressed, but other authors have also found high densities of T. fluviatilis in some karstic springs (e.g., Barquín and Death, 2004, 2006). Whereas insects generally predominate in streams, karstic springs are generally dominated by invertebrates with high calcium requirements, such as crustaceans (amphipods, isopods) and gastropods (Barquín & Death, 2004).

Our main aim was to determine the reasons for the decreased densities of T. fluviatilis downstream from the springs. Therefore, we first investigated density variations within a karstic spring. We then quantified changes in densities of T. fluviatilis downstream from the spring in relation to changes in the environmental conditions. Given the importance of CaCO₃ for shell building (Dahm et al., 1998; Allan & Castillo, 2007), our hypothesis was that alkalinity was the main driver of the abundance of this species. Third, given the high thermal stability of springs, we hypothesised that life-history parameters would be little affected by yearly temperature-related events and that reproduction should therefore occur year-round.

MATERIALS AND METHODS

Study area

Our study was conducted at the spring source of the Anços River (Olhos de Água; N 39°58’43″, W 08°34’21″; 69 m asl), which is fed by two major groundwater outlets and has a mean yearly flow of 50 000 m³ (Cunha, 1990). The spring is bordered essentially by Salix spp. and Populus spp. and is located in an agricultural area. The streambed downstream from the spring includes stony, sandy and muddy areas where Apium nodiflorum thrives. The area of the spring is ~ 20 × 12 m, and the width of the river ranges from ~ 15 m just outside of the spring to ~ 30 m at its widest portion in the study area.

Spatial distribution

To understand the spatial distribution of T. fluviatilis, i.e., if the species is associated with a particular type of substrate and other conditions such as depth and current velocity, we randomly collected 31 Surber samples (0.3 × 0.3 m
opening and 0.5 mm mesh size) from the spring in December 2005. The samples were distributed across three distinct areas: (A) the upstream outlets of the groundwater, which feed a small pond (n = 9); (B) a very wide area farther downstream, where macrophytes usually grow (n = 20); and (C) lateral smaller outlets discharging into the wide area (n = 3). At each sampling point (Surber), we assessed the depth, current (VALEPORT 15277), the dominant substrate size (boulder: > 256 mm; cobble: 64–256 mm; pebble/gravel: 2–64 mm; sand: 0.06–2 mm; and silt < 0.06 mm) and the presence of macrophytes. We also measured the temperature, pH, electrical conductivity and dissolved oxygen with field probes. The biological samples were preserved in 4 % formalin (neutralised with sodium tetraborate) and transported to the laboratory. In the laboratory, the samples were washed with running tap water over a column of sieves (0.5, 1 and 2 mm), and the specimens of *T. fluviatilis* were sorted and counted.

**Longitudinal distribution**

We investigated the occurrence of the population of *T. fluviatilis* at different distances beyond the spring, and we determined the relationship between these population densities and changes in the physical and chemical conditions. Three Surber samples were taken at 0, 25, 100, 200, 800, 1000, 1800 and 3800 m downstream from the spring in three different habitats (macrophyte beds, stony substrates and sandy substrates) in June and July 2006 and processed as described above. Simultaneously, we measured the physical and chemical variables listed above and the type of flow (riffle, run, pool). We also took one L of water in a clean plastic bottle for ion determinations (F⁻, CH₃CO₂⁻, Cl⁻, NO₂⁻, NO₃⁻, SO₄²⁻, Na⁺, NH₄⁺, K⁺, Mg²⁺, Ca²⁺; ion chromatograph, Dionex DX-120, Sunnyvale, CA). Soluble reactive phosphorus (SRP) was determined by the ascorbic acid method and alkalinity was determined by titration with H₂SO₄ to a final pH of 4.5 (APHA, 1995).

**Population dynamics**

In karstic springs, thermal conditions tend to be buffered against seasonal variations during the year. Because reproduction in invertebrates depends strongly on temperature, we hypothesised that in such environments, seasonality in reproduction should be less apparent. To test this hypothesis, we sampled the population monthly from June 2006 to June 2007 in three habitats: macrophyte beds, stony substrates and sandy substrates. Samples were taken with a hand net (0.25 × 0.25 m and 0.5 mm mesh size) covering an area comparable to the Surber sampler (0.3 × 0.3 m). Stones were placed in a bucket and washed with a brush. In the laboratory, the shell length (L) of each individual was measured to ± 0.1 mm to allow the calculation of individual biomass (B, as ash-free dry mass-AFDM in mg; $B = 0.17 e^{0.5361L}$, $R^2 = 0.80; p < 0.001$; length ranged from 1.4 to 10.6 mm).

**Data treatment**

**Spatial distribution**

To understand which variables best explained the distribution of *T. fluviatilis*, we performed a principal components analysis (PCA) on the environmental variables pH, temperature, dissolved oxygen, electrical conductivity, current velocity, depth and substrate type (normalised; PRIMER 6). The densities of *T. fluviatilis* were log (x + 1) transformed and regressed against PCA axes 1 and 2. A multiple regression analysis (backwards; STATISTICA 6) was also used to investigate the importance of each environmental variable for the numbers of *T. fluviatilis* (log (x + 1)). Finally, to assess the importance of substrate in the distribution of *T. fluviatilis*, we compared the abundance of the species in three habitats (stony substrates, sandy substrates and macrophyte beds) in spring area B by an ANOVA. To guarantee the normal distribution of data and the homogeneity of variances, we used a
Shapiro-Wilk W test (STATISTICA 6) and a Bartlett Chi-Squared test (STATISTICA 6).

Longitudinal distribution

The abundance data were subjected to a two-way analysis of variance (2-way ANOVA), with distance to the source and habitat type as factors (STATISTICA 6). The environmental data were again subjected to a PCA and the density data were regressed against PCA axes 1 and 2 (PRIMER 6).

Population dynamics

Population dynamics was analysed in terms of biomass (B, in mg). The densities \((\log(x + 1))\) were higher in stony substrates than in sand and macrophytes (2-way ANOVA) and we therefore further investigated changes in densities \((\log(x + 1))\) and biomass only in stony substrates across the year with a 1-way ANOVA (STATISTICA 6), with time as a factor. The growth rates and secondary production were subjected to a modal analysis in ANAMOD software (Nogueira, 1992). Because no cohorts were identified, we calculated secondary production from July 2006 to June 2007 by the Hynes size-frequency method, according to Krueger and Martin (1980), in which the annual production \((P)\) is given by:

\[
P = \left\{ \sum_{j=1}^{i} \left( \bar{n}_j - \bar{n}_{j+1} \right) \cdot \left( W_j \cdot W_{j+1} \right)^{1/2} \right\} \cdot \frac{365}{\text{CPI}}
\]

in which \(i\) is the number of size classes, \(n_j\) is the number of individuals reaching a particular size \(j\) in one year, \(W_j\) is the mean individual biomass in size \(j\) and CPI (cohort production interval) is the mean number of days that an individual in a cohort takes to reach the maximum size (3.5 years * 365 days; Kirkegaard, 2006). We then calculated \(P/B\). We recognise that the CPI in Kirkegaard (2006) was determined for a different environment, in which the temperatures were lower than those reported in Ançóes spring. In this warmer environment, it is predictable that \(T.\ fluvialis\) would have shorter life cycles. Therefore, the real CPI would be lower than that determined from the study of Esrom Lake by Kirkegaard (2006) and the production estimate should be considered a minimum value.

RESULTS

Spatial distribution

The three spring locations, A, B and C, did not differ markedly in most of their environmental variables (Table 1). Nevertheless, they were segregated by PCA, with axes 1 and 2 explaining 32 and 23 % of the variation, respectively (Fig. 1). The abundance of \(T.\ fluvialis\) was negatively correlated with PCA axis 2 according to the equation

\[
\log(\text{Abundance} + 1) = -0.48 \times (\text{PCA2}) + 1.05 (\pm 0.12)
\]

\(R^2 = 0.59\); \(p < 0.001\); \(n = 26\).

Figure 1. Principal Components Analysis of 26 samples \(\times\) 10 environmental variables for Ançóes spring. Environmental variables: pH, temperature, dissolved oxygen, electrical conductivity, current velocity, depth and substrate type (according to their predominance) at several locations within the spring (see Table 1 for locations A, B and C). Axes 1 and 2 explained 32 % and 23 %, respectively, of the variation. Análisis de Componentes Principales de 26 muestras \(\times\) 10 variables ambientales en el manantial del río Ançóes: pH, temperatura, oxígeno disuelto, conductividad eléctrica, velocidad de la corriente, profundidad, y tipo de sustrato (de acuerdo con su predominancia) en varios puntos dentro del manantial (ver Tabla 1 para puntos A, B, y C). Ejes 1 y 2 explicaron 32 % y 23 % de la variación.
Table 1. Physical and chemical variables measured in situ in an upstream sector of the spring (A), lateral entrances of groundwater (C) and the large lower section where the waters meet (B). Mean (±1 SE) [minimum; maximum]; n = 23. Variables fisicoquímicas medidas in situ aguas arriba del manantial (A), entradas laterales de aguas subterráneas (C), y aguas abajo donde confluyen las aguas de los puntos anteriores (B). Media y Error Estándar (±1 SE) [mínimo; máximo]; n = 23.

<table>
<thead>
<tr>
<th>Physical and chemical variables</th>
<th>Upstream (A)</th>
<th>Downstream (B)</th>
<th>Lateral (C)</th>
</tr>
</thead>
<tbody>
<tr>
<td>pH</td>
<td>7.1 (± 0.1)</td>
<td>7.2 (± 0.1)</td>
<td>7.1 (± 0.1)</td>
</tr>
<tr>
<td></td>
<td>[6.9; 7.2]</td>
<td>[7.0; 7.4]</td>
<td>[7.0; 7.1]</td>
</tr>
<tr>
<td>Temperature (°C)</td>
<td>16.0 (± 0.1)</td>
<td>16.1 (± 0.1)</td>
<td>16.3 (± 0.3)</td>
</tr>
<tr>
<td></td>
<td>[15.9; 16.1]</td>
<td>[16.0; 16.5]</td>
<td>[16.0; 16.5]</td>
</tr>
<tr>
<td>Oxygen saturation (%)</td>
<td>105 (± 2)</td>
<td>112 (± 3)</td>
<td>101 (± 0.00)</td>
</tr>
<tr>
<td></td>
<td>[102; 108]</td>
<td>[105; 114]</td>
<td></td>
</tr>
<tr>
<td>Dissolved oxygen (mg l⁻¹)</td>
<td>10.3 (± 0.2)</td>
<td>10.9 (± 0.3)</td>
<td>9.9 (± 0.00)</td>
</tr>
<tr>
<td></td>
<td>[10.1; 10.6]</td>
<td>[10.0; 11.2]</td>
<td></td>
</tr>
<tr>
<td>Conductivity (µS cm⁻¹)</td>
<td>483 (± 0.00)</td>
<td>487 (± 5)</td>
<td>500 (± 12)</td>
</tr>
<tr>
<td></td>
<td>[485; 504]</td>
<td>[486; 507]</td>
<td></td>
</tr>
<tr>
<td>Velocity (m s⁻¹)</td>
<td>0.4 (± 0.3)</td>
<td>1.3 (± 0.6)</td>
<td>0.3 (± 0.2)</td>
</tr>
<tr>
<td></td>
<td>[0.1; 0.9]</td>
<td>[0.3; 2.7]</td>
<td>[0.0; 0.5]</td>
</tr>
<tr>
<td>Depth (cm)</td>
<td>52 (± 15.3)</td>
<td>22 (± 6.8)</td>
<td>31 (± 5.5)</td>
</tr>
<tr>
<td></td>
<td>[40; 85]</td>
<td>[12; 36]</td>
<td>[25; 35]</td>
</tr>
<tr>
<td>Boulder substrate (%)</td>
<td>16 (± 33)</td>
<td>32 (± 36)</td>
<td>33 (± 12)</td>
</tr>
<tr>
<td></td>
<td>[0; 100]</td>
<td>[0; 90]</td>
<td>[20; 40]</td>
</tr>
<tr>
<td>Cobble substrate (%)</td>
<td>7 (± 12.5)</td>
<td>8 (± 9.6)</td>
<td>0 (± 0.0)</td>
</tr>
<tr>
<td></td>
<td>[0; 40]</td>
<td>[0; 30]</td>
<td></td>
</tr>
<tr>
<td>Pebble substrate (%)</td>
<td>19 (± 22)</td>
<td>22 (± 28)</td>
<td>28 (± 18)</td>
</tr>
<tr>
<td></td>
<td>[0; 60]</td>
<td>[0; 90]</td>
<td>[10; 45]</td>
</tr>
<tr>
<td>Sandy substrate (%)</td>
<td>26 (± 31)</td>
<td>21 (± 34)</td>
<td>35 (± 15)</td>
</tr>
<tr>
<td></td>
<td>[0; 80]</td>
<td>[0; 100]</td>
<td>[20; 50]</td>
</tr>
</tbody>
</table>

This equation suggests that high current velocity, a substrate larger than sand, high electrical conductivity, shallow areas and high temperature were all important factors affecting the numbers of *T. fluviatilis*. Indeed, the numbers of specimens were explained (multiple regression analysis) by three factors: the substrate type (with predominance of boulders or cobbles), depth and electrical conductivity by the equation:

\[
\log (\text{Abundance} + 1) = 0.02 \times \text{boulders} - 0.034 \times \text{depth} + 0.048 \times \text{conductivity} + 0.03 \times \text{cobbles} - 21.84( \pm 0.378) \]

\[ R^2 = 0.86; \ p < 0.001; \ n = 26. \]

The samples of *T. fluviatilis* in the wide B area of the spring were taken in three different habitat types: stony substrates, sandy substrates and macrophyte beds. The densities of *T. fluviatilis* in sandy substrates (monthly average) reached 100 individuals m⁻²; this value was 2.5 times higher in macrophyte beds (250) and almost 100-fold higher in stony substrates (8300; ANOVA by ranks; \( H = 13.814; \ p < 0.001; \) Fig. 2).

**Longitudinal distribution**

The densities of *T. fluviatilis* were highest at the spring and 200 m downstream. They then decreased longitudinally. At 3800 m downstream,
we found no specimens of *T. fluviatilis* in the samples (Fig. 3). Oxygen, Na\(^+\), K\(^+\) and Cl\(^-\) tended to increase downstream (Table 2). The two main axes of the PCA on the abiotic factors

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**Figure 2.** *Theodoxus fluviatilis* abundances across three substrate types sampled in the Anços spring in June and July 2006. ANOVA by ranks (\(n = 18\)); different letters denote significant differences according to a post-hoc (Tukey) test. Mean ± 1 SE.

**Table 2.** *Theodoxus fluviatilis* mean abundance and physical and chemical variables measured at the Anços spring and downstream. *Abundancias medias de Theodoxus fluviatilis y variables fisicoquímicas medidas en el manantial del río Anços y tramos aguas abajo.*
explained 80% of the variation (PCA1 = 63%; PCA2 = 17%; Fig. 4). The PCA biplot did not display the sites in their sequence along the river. The abundances of *T. fluviatilis* were not significantly correlated with any axis (Spearman rank correlation; *p* > 0.05) or with any measured environmental variable (*p* > 0.05). However, the physical and chemical changes across the stretch studied were very small (Table 2). We were therefore unable to relate the variation in *T. fluviatilis* densities to variations in the abiotic environment.

### Population dynamics

The habitat type and month of the year significantly affected the densities of *T. fluviatilis*. The highest densities were found on stony substrates (2-way ANOVA, *F* = 1171.219, *p* < 0.001; date *p* < 0.009; habitat *p* < 0.001, date × habitat *p* < 0.005). For stony substrates, the highest densities were observed in August and September (20 000 to 24 000 individuals m⁻²) and the lowest densities in December and May (~ 3000 individuals m⁻²), but no differences were observed (1-way ANOVA, *F* = 1.954, *p* = 0.082, d.f. = 11; Fig. 5). Floods in October prevented sampling during that month. The biomass varied less than the densities, with values generally ranging from 7 to 27 g AFDM m⁻², but the monthly differences were not significant because of high variation and differences in dominant size classes during the year (1-way ANOVA, *F* = 2.082, *p* = 0.064, d.f. = 11; Fig. 6).

Egg masses were observed on the substrates during the entire year. Smaller individuals predominated from June through August 2006 and again from May 2007 onwards (Fig. 7); however, small individuals were always present. Taken together, this information suggests continuous reproduction throughout the year. Cohorts were difficult to identify. However, recruitment was likely more intense in summer because peaks of density occurred from July through September 2006 and again in June 2007 and during this period there was a predominance of small individuals. During the year, the overall most abundant size class was 1.13–1.34 mm, the estimated
Figure 7. Size-class frequencies of *Theodoxus fluviatilis* in Anços spring from June 2006 through June 2007. *Frecuencia de clases de tamaño de Theodoxus fluviatilis en el manantial del río Anços entre junio de 2006 y julio de 2007.*
Table 3. Estimated *Theodoxus fluviatilis* population parameters at the Anços spring, using the size frequency method. $P =$ production; $B =$ biomass. Parámetros poblacionales estimados para *Theodoxus fluviatilis* en el manantial del río Anços, de acuerdo con el método de clases de frecuencia. $P =$ producción; $B =$ biomasa.

<table>
<thead>
<tr>
<th>Parameters</th>
<th>Values</th>
</tr>
</thead>
<tbody>
<tr>
<td>P cohort (hypothetical)</td>
<td>90.40 g AFDM m$^{-2}$</td>
</tr>
<tr>
<td>P annual</td>
<td>25.83 g AFDM m$^{-2}$ year$^{-1}$</td>
</tr>
<tr>
<td>Mean B</td>
<td>15.52 g AFDM m$^{-2}$</td>
</tr>
<tr>
<td>P cohort/Mean B</td>
<td>5.826</td>
</tr>
<tr>
<td>P annual/Mean B</td>
<td>1.665 year$^{-1}$</td>
</tr>
</tbody>
</table>

Mean biomass was 15.52 g AFDM m$^{-2}$, the annual production was 25.83 g AFDM m$^{-2}$ and the $P/B = 1.665$ year$^{-1}$ (Table 3). The temperatures at the Anços spring during the study period ranged from 15.3 to 16.6 °C, with minimal daily and yearly variation compared with temperatures recorded 900 m downstream, where temperatures in the warmest months ranged from 16.0 to 21.4 °C with maximum daily variations from 3 (shade) to 5 °C (river exposed to sun; Fig. 8).

**DISCUSSION**

*T. fluviatilis* congregated preferentially on coarse substrates, primarily cobbles and boulders, in shallow areas with a rapid current. Lithophilic congregation was reported previously by other authors (Fertter & Graham, 1978; Bunje, 2005; Kirkegaard, 2006) and may be related to the greater stability of coarse sediments. Current and substrate particle size are most likely related, with fine particles associated with deposition areas and coarse particles with erosion areas (Giller & Malmqvist, 1998). Stable and shallow sediments in fast-flowing waters favour the establishment of the periphytic community (Poff et al., 1990; Poff & Ward, 1992, 1995; Cardinale et al., 2002; Opsahl et al., 2003) where *T. fluviatilis* feeds. Indeed, this species has been reported as a scraper, feeding mainly on diatoms (Zettler et al., 2004; Kirkegaard, 2006).

We also observed *T. fluviatilis* associated with macrophytes, as reported previously by Carlsson (2000) and Zettler et al. (2004), and we presume that the reason is the same: substrate for periphyton and high primary productivity. In-

![Figure 8](image-url)
Indeed, macrophyte production was very high, as judged by the vigorous growth of *A. nodiflorum*. Therefore, the productivity of the periphyton was likely similarly high given the light availability, although this parameter was not measured.

After examining the spatial distribution in the spring, we investigated the longitudinal distribution. We found no strong relationship between any of the measured chemical and physical variables and the numbers of individuals. This lack of relationship was not surprising in terms of water chemistry because the chemical changes were too small to be ecologically relevant. In terms of physical conditions, although sandy substrates were more abundant downstream, stony structures were always present. Moreover, macrophytes (substrates for *T. fluviatilis*) were present along the entire section studied and the numbers of individuals on this substrate decreased farther downstream. These findings are consistent with other studies reporting little change in water chemistry from springs to downstream reaches (e.g., Barquin & Death 2011).

We should therefore consider other reasons for changes in *T. fluviatilis* abundance downstream from the karstic spring. One of the explanations could be the stability of conditions in the spring, which would buffer the typical stochasticity of rivers and favour the role of biotic interactions in maintaining the community structure. Therefore, the long-living gastropods could likely outcompete other periphytic consumers, such as baetid mayflies (but not gammarids).

Another explanation for the high abundance of *T. fluviatilis* in the spring, with decreased abundance downstream, could be continuous recruitment in the spring, favoured by constant and favourable temperatures. Temperature is known to control reproductive activity and growth in *T. fluviatilis* (Fretter & Graham, 1978; Kirkegaard, 2006). Indeed, we found that the temperature was fairly constant at approximately 16°C, and egg capsules were observed year-round on stones and other snail shells, suggesting continuous reproduction. Moreover, in addition to the direct effect of temperature on *T. fluviatilis*, light availability affects primary production and therefore energy resources for the population. The observation of continuous recruitment contrasts with the pattern reported for populations in a Danish lake, in which the two periods of maximum recruitment that occurred were related to temperature (Kirkegaard, 2006).

The secondary production in the spring was calculated as 25.8 g AFDM m⁻² year⁻¹ and P/B 1.67 year⁻¹. The values estimated by Kirkegaard (2006) in a Danish lake were approximately 10-fold lower (1.8–2.3 g AFDM m⁻² year⁻¹). Our mean density value was 7-fold higher than that in the Danish site (8300 vs. 1200 individuals m⁻²). These values suggest favourable conditions for *T. fluviatilis* in the Anços spring.

**CONCLUSION**

We suggest that the high densities of *T. fluviatilis* in the Anços karstic spring, and presumably in other karstic springs in central Portugal, are the result of high and stable temperatures, which favour continuous reproduction. Lower winter temperatures farther downstream do not eliminate the species but will restrict its reproduction and therefore its densities. The factors affecting the distribution of *T. fluviatilis* will likely affect the distribution of other species. For example, *Echinogammarus meridionalis* is another invertebrate occurring in very high numbers in the spring. The abundance of this species decreases downstream (although not as dramatically as *T. fluviatilis*).

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